

### Research article

# Above- and belowground traits along a stress gradient: trade-off or not?

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The role of plant traits in shaping community assembly along environmental gradients is a topic of ongoing research. It is well accepted that plant traits of aboveground organs tend to be conservative in stressful conditions. However, there is limited understanding of how belowground traits respond. Plants may have similar strategies above and belowground, but an intriguing possibility is that there is a tradeoff between above and belowground traits of communities to both ensure efficient resource-use and limit niche overlap along the gradient. To test this, we asked whether the response of above and belowground traits of communities is coordinated or not along a stress gradient in Mediterranean sand dune communities. We analyzed 80 vegetation plots in central Italy to test for coordinated vs independent patterns in above vs belowground plant traits using community weighted mean and standardized effect size of functional richness. Our results show that plant communities close to the sea, which experience higher stress, were characterized by higher convergence towards aboveground resource conservation and conservative wateruse strategies but belowground resource acquisition, consistent with a strong effect of habitat filtering and an above-belowground tradeoff favoring adaptation to harsh and dry conditions. At the opposite end of the gradient with lower stress, plants exhibited higher trait diversity for both above and belowground traits, but overall a dominance of aboveground fast resource acquisition and generally acquisitive water-use strategies, combined with conservative belowground strategies. This suggests that fast growth rate aboveground was compensated by more conservative fine-root strategies, but processes such as competition limited niche overlap overall. Our findings provide new insights into the relationship between functional traits and environmental gradients in plant communities, shedding light on the tradeoffs between the above and belowground dimensions.

Keywords: belowground traits, coastal dunes, competition, habitat filtering, root traits, stress

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#### Introduction

Understanding the mechanisms shaping plant communities along environmental gradients is a long-standing ecological issue (Funk et al. 2017). In the last decades, researchers shed light on process contributing to community assembly by studying species' functional traits (Götzenberger et al. 2012, Funk et al. 2017, Laliberté 2017, de Bello et al. 2021, Bricca et al. 2023a). Nowadays, it is widely accepted that a local plant communities can originate from random processes such as dispersal events or priority effects or by deterministic processes operating on species niches whose relative contribution depends on the harshness of local environmental conditions (Götzenberger et al. 2012, Coyle et al. 2014, Bricca et al. 2022). For example, stressful conditions select species with largely overlapping niches that are functionally convergent (i.e. sharing more similar trait values than expected by chance) because habitat filtering processes reduce the range of successful strategies within the plant community (Weiher and Keddy 1995, Cornwell et al. 2006). On the contrary, under less stressful environmental conditions, niche partitioning processes limit the similarity of coexisting species (limiting similarity), leading to a wider range of successful strategies than expected by chance (functional divergence; MacArthur and Levins 1967). Current knowledge suggests that both random and deterministic processes can occur simultaneously, potentially affecting different sets of species traits (Carboni et al. 2014, Bricca et al. 2021).

Most investigations addressing assembly rules have focused on few aboveground traits, namely specific leaf area (SLA), plant height (H), and seed mass (SM), jointly referred to as the LHS scheme of Westoby (1998) (Lalibertè 2017). These traits, collectively considered representative of the aboveground functional dimension, relate to resource exploitation (SLA), competitive ability for light (H), and reproductive ability (SM; Díaz et al. 2016). However, using only LHS traits leaves a substantial part of functional space largely unexplored, in particular the belowground dimension (Carmona et al. 2021, Klimešová and Herben 2023). Recently, belowground plant ecology research is experiencing a spiking interest, with advancements focused on fineroot traits reflecting shifts from acquisitive to conservative strategies (specific root length, SRL; root tissue density, RTD) and on belowground organ traits expressing wateruse strategies (belowground organ water content, BOWC) (Freschet et al. 2010, Stanisci et al. 2010, de Bello et al. 2012, Montagnoli et al. 2012, 2014, Kramer-Walter et al. 2016, de la Riva et al. 2016, 2019, Laliberté 2017, Erktan et al. 2018, Carmona et al. 2021, Laughlin et al. 2021). These studies describe how belowground traits coordinate with aboveground ones to determine viable species ecological strategies, yet they do not provide information on community assembly mechanisms. At the same time, few efforts have been made to investigate how belowground traits contribute to plant community organization along stress gradient (Holdaway et al. 2011, Spitzer et al. 2022), and attempts

to examine aboveground and belowground plant traits in a single framework are still rare.

The simplest hypothesis is that plants may have similar strategies above and belowground, but an intriguing possibility is that there is a tradeoff between above and belowground traits of communities. For example, it has been proposed that increasing stress promotes a shift in competition from aboveground to belowground resources (Reynolds and D'antonio 1996, Coomes and Grubb 2000, Rajaniemi 2003, de Bello et al. 2012, Tilman 2020). In particular, the 'resource allocation theory' builds on the idea that under favorable conditions plant species show an aboveground acquisitive strategy by allocating relatively more carbon to aboveground (fast-growing strategies), rather than belowground organs, because of the higher competition for light (Tilman 2020). Conversely, under stressful conditions, species will shift towards belowground acquisitive strategies, limiting the aboveground biomass investment (slow-growing strategies) due to intense competition for water and soil nutrients (de Bello et al. 2012). This variation in the mode of competition might also be mirrored in opposite aboveground and belowground functional diversity patterns. In line with the limiting similarity hypothesis, we may expect that when species show a high degree of niche overlap for certain aboveground traits, the belowground functional dimensions should ensure sufficient niche partitions to allow coexistence. For example, it has been observed that along a large-scale gradient of productivity, clonal and bud bank traits diverged while LHS strategies tended to converge (Vojtkó et al. 2017). Still, opposite functional diversity patterns between aboveground and belowground ecological niche might not be the only possible coexistence solution for plants. Recently, a global investigation at the species level highlighted how LHS and fine-root traits are not aligned along the same functional axes, thus suggesting an independent contribution of these functional dimensions (Carmona et al. 2021). Nonetheless, an independence of aboveground and belowground plant traits ruling the community assembly process has never been argued before and evidence is still lacking.

In consideration of the above, this study aims at shedding light on assembly patterns of above- and belowground traits in sand dune plant communities, which are known to be highly stressful environments. Indeed, due to the presence of a steep sea-inland gradient running perpendicular to the coastline, these ecosystems feature harsh abiotic conditions, whose intensity decreases towards the inland, arguably in parallel to increasing intensity of competition (Santoro et al. 2012; Fig. 1, Supporting information). This gradient is responsible for a considerable turnover in species composition and functioning over very short distances, making coastal dunes an ideal system to test the effect of variation in stressful conditions on plant communities (Acosta et al. 2009, Carboni et al. 2011, 2016, Ciccarelli and Bona 2022). Specifically, plant communities close to the sea, which experience more stressful conditions (e.g. foredunes dominated by perennial species) are typically characterized by species with taller size and aboveground slower-growing strategies (e.g.

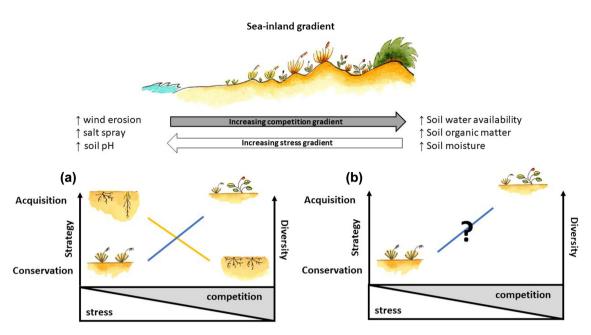


Figure 1. Conceptual framework highlighting different kinds of relationships between aboveground and belowground traits in relation to the sea-inland gradient as proxy of stress gradient; (a) according to the resource allocation theory and limiting similarity hypothesis, above-and belowground traits have opposite resource exploitation strategies (acquisition versus conservation) and trait diversity patterns; (b) alternatively, above- and belowground traits resource exploitation strategies and trait diversity are not negatively coordinated. Aboveground and belowground variation is represented by blue and orange lines, respectively. Variation of abiotic conditions along sea—inland gradient according to Carboni et al. (2011, Supporting information).

Calamagrostis arenaria subsp. arundinacea and Thinopyrum junceum). On the contrary, plant communities away from the sea in less stressful conditions (e.g. back dunes dominates by annual species) showed species with shorter size and aboveground faster-growing strategies (e.g. Silene canescens, Festuca fasciculata) (Acosta et al. 2006, Ricotta et al. 2012, Ciccarelli 2015, Carboni et al. 2016, Ciccarelli and Bona 2022). However, how belowground traits respond is still unknow. Thus, we selected a set of six functional traits capturing the largest amount of variation in above- (LHS) and below-ground functional dimensions (SRL, RTD, BOWC; Carmona et al. 2021, Lubbe et al. 2021). We investigated if above- and below-ground traits contribute to community assembly along this stress gradient, similarly, showing opposite patterns or behaving independently. We mainly hypothesize two alternative scenarios (Fig. 1):

1) Patterns in aboveground traits will shift from dominance of conservative strategies (low SLA, high H, high SM) and/or functional convergence due to habitat filtering under the more stressful conditions close to the coastline, to dominance of acquisitive strategies (high SLA, low H, small SM) and/or functional divergence due to competition and limiting similarity under the milder conditions of the inland (Fig. 1a). In turn, the belowground functional pattern should shift from the dominance of acquisitive belowground traits (high SRL, low RTD, high BOWC) and/or functional divergence under stressful conditions (to compensate for high similarity in aboveground traits), to the dominance of conservative belowground strategies

- (low SRL, high RTD, low BOWC) and/or functional convergence under less stressful conditions (Fig. 1a);
- 2) Aboveground traits changes as in the previous scenario, whereas belowground traits are expected not to be negatively coordinated. For example, belowground traits could:

  1) change similarly to aboveground traits or 2) show no variation at all due to either the constant influence of a certain driver (i.e. habitat filtering or limiting similarities) along the whole gradient or because of random processes (Fig. 1b).

#### Material and methods

#### Study area

We studied recent (Holocene) coastal dunes occupying a narrow strip (<150 m) of about 16 km length along the seashore of the Tyrrhenian coast of central Italy (Lazio Region). The area has a Mediterranean climate characterized by a long dry summer with low annual rainfall (Acosta et al. 2003). In coastal dunes, vegetation is generally organized along a seainland gradient that runs perpendicular to the coastline and gives rise to a well-known 'coastal zonation' (Van Der Maarel 2003, Acosta et al. 2007). Moving away from the shoreline, upper beach communities colonizing the drift line are followed by embryonic dunes, mobile dunes and a stabilized zone dominated by herbaceous plant communities (so called 'dune grasslands'), ending up in the Mediterranean maquis (Acosta et al. 2009).

#### Above- and belowground traits sampling

From the 'RanVegDunes' database (Sperandii et al. 2017), we extracted 80 georeferenced random plots (4 m²) according to the following criteria: 1) plot classified as herbaceous habitat; 2) plot not invaded by alien species; 3) plot where aboveground traits data were available at least for those species whose relative cumulative cover reached 80% of plot total cover (separately for all the three aboveground traits; Pakeman and Quested 2007); 4) plots where aboveground traits data were available at least for two species. According to the Annex I of the Directive 92/43/EU, these plots included three different habitats: embryonic dunes (code 2110), mobile dunes (code 2120), and dune grasslands (code 2230). For each plot we obtained the distance from the sea, which has been demonstrated to be a good proxy of the sea-inland gradient, with habitats close to sea characterized by higher stress conditions (lower soil moisture and organic matter, higher pH, salt spray and wind erosion, Fig. 1, Carboni et al. 2011, Bazzichetto et al. 2016). Aboveground traits, i.e. plant height, SLA and seed mass, were extracted from 'TraitDunes' (TRY database no. 429; Kattge et al. 2020), a database including plant traits measured on coastal dune species in central Italy. Belowground traits were measured in the field for a total of 16 species namely dominant species for which aboveground trait data were available. Trait sampling took place in 2018, during the spring season. For each of the 16 species, we dug out 3-5 individuals using a little shovel and hands and, for each individual, we measured the water content of the main belowground organ (BOWC; %), e.g. bulb, tap root, tuber, rhizome, which is informative on water-use strategies, and two fine-root traits, i.e. specific root length (SRL; cm mg<sup>-1</sup>) and root tissue density (RTD; mg cm<sup>-3</sup>), capturing the acquisitive and conservative strategies respectively (de Bello et al. 2012, de la Riva et al. 2019). Specifically, we collected individuals at least 10 m from each other to avoid sampling the same individuals for clonal species. Individuals collected in the field were placed in plastic bags with water and stored in a cold dark box. Once in the laboratory, individuals within plastic bags were stored in the fridge. To measure BOWC, we followed the approach of de Bello et al. (2012): within the following 24 h we cut and selected as sample the oldest parts of the main belowground organs (which in case of rhizomes correspond to the part closer to the shoot; Bricca et al. 2020), as this trait is related to resource conservation. Water-satured samples of belowground organs were weighted and then oven-dried at 60°C for approximately 96 h, i.e. until they were completely dry. Finally, similarly to de Bello et al. (2012) we expressed the BOWC as the percentage of water content (i.e. as one minus the ratio of dry weight to fresh water-saturated weight). Although the relevance of this trait remains poorly assessed to date, its use in combination with leaf and fine-root traits is promising for understanding water use patterns (de Bello et al. 2012, Lubbe et al. 2021). Concerning fine-root traits, within the following 24 h in the laboratory we gently cleaned fineroots with water and stored them in a 50% ethanol solution to keep them fresh for longer periods of time. Then, according to the global protocol of trait collecting and measuring (Perez-Harguindeguy et al. 2013) fine-root samples were first scanned in a small amount of water at a resolution of 800 dpi, and then dried (at least 48 h at 60°C) and weighed. Scanned fine-roots images were processed with the WinRHIZO software (Regent Instruments). To obtain fine-roots mainly involved in absorption, we used samples with diameters less than 1 mm (De Battisti et al. 2020). Specifically, the average root diameters across all our samples were < 0.5 mm, with the only exception of *Pancratium maritimum* whose average root diameter was 2.27 mm. A description of each trait is provided in Table 1. Finally, for each species we also recorded the annual lifespan (AL) at the whole plant level as a categorical trait with values of 1 and 0 for annual versus not annual species, respectively.

Above- and belowground traits values were  $\log_{10}$ -transformed before performing data analysis to improve the normality of trait distribution, except the BOWC since it was not necessary (Májeková et al. 2016).

#### Statistical analysis

#### Community analysis

To quantify variation in the functional strategies at community level, we calculated the community weighted mean (CWM; Garnier et al. 2004) for each trait separately:

$$CWM = \sum_{i=1}^{S} p_i x_i \tag{1}$$

where S is the number of species,  $p_i$  is the relative cover of species i (i = 1, 2, ..., S), and  $x_i$  is the trait value for species i. CWMs provide complementary information to the functional diversity index, quantifying variation in the dominant functional strategies in relation to the environmental gradient. Variation of CWMs in relation to sea distance was analyzed using a fourth corner analysis (Zelený 2018). This method explores the mean trait responses to environmental gradients by simultaneously analysing the information contained in three matrices: a matrix of species composition (L, abundances of species in samples), a matrix of sample attributes (e, sea distance), and a matrix of species attributes (t, species traits). The fourth-corner approach combines all three matrices (e, t and L) by calculating the (weighted) correlation between standardized species and sample attributes using a species composition table as the weight matrix (ter Braak et al. 2018). In practice, the fourth-corner correlation cannot reach -1 or 1 unless species distributions are perfectly ordered across columns and rows (Peres-Neto et al. 2017). Together with the fourth corner we selected the row-based permutation algorithm which tests the link between plot attributes (e.g. CWM) and species composition, and it is performed by randomly shuffling plot attributes 999 times across plots. A p-value ( $p_r$ , row-based permutation) is estimated as the number (+1) of random squared correlations equal to or greater than the observed squared value divided by the number of permutations +1. By adding 1, the observed values are included as one possible value under the null distribution

Table 1. List of the aboveground and belowground traits considered, their codes (unit of measurement), their function, their expected ecological strategies in sandy dunes plant communities. Shift of trait values underlying variation in the resource strategies with increase of stress are reported with an arrow facing down if a negative relationship is expected, and facing up if a positive relationship is expected.

Functional dimension	Trait (code, unit)	Function	Resource strategies and trait states (source)	Distance to sea
Life history	Annual lifespan (AL)	Persistence	Acquisitive – annual Conservative – non annual	1
	Plant height (H, m)	Competitive ability	(Perez-Harguindeguy et al. 2013) Conservative – high H Acquisitive – low H	<b>↓</b>
Aboveground	Specific leaf area (SLA, mm² g <sup>-1</sup> )	Photosynthetic activity	(Acosta et al. 2006) Acquisitive – high SLA Conservative – low SLA	<b>↑</b>
	Seed mass (SM, g)	Dispersion and reproduction	(Acosta et al. 2006) Conservative – high SM Acquisitive – low SM	Ţ
	Specific root length (SRL, cm mg <sup>-1</sup> )	Resource acquisition	(Carboni et al. 2016) Acquisitive – high SRL Conservative – low SRL	$\downarrow$
Belowground	Root tissue density (RTD, mg cm <sup>-3</sup> )	Resource transport	(Kramer-Walter et al. 2016) Conservative – high RTD Acquisitive – low RTD	<b>↑</b>
	Belowground organ water content (BOWC, %)	Water-use	(Kramer-Walter et al. 2016) Acquisitive – high BOWC Conservative – low BOWC (Freschet et al. 2010)	1

(Peres-Neto et al. 2017). In addition, as CWM incorporates the cover of each species, we also calculated an unweighted version of this index, the community mean (CM), using only presence/absence data. In this way we were able 1) to disentangle if community variation depends on species turnover (i.e. changes species identity across plots) or on the variation of the cover of few dominant species across plots having same species identity (de Bello et al. 2021); 2) avoid a potential mismatch for indices which combine aboveground species cover data with belowground trait values (Tsakalos et al. 2022). To analyze CMs, we used the fourth corner analysis and the same modified permutation test as for CWMs. We concluded on the presence of tradeoffs if plant communities showed contrasting aboveground and belowground strategies (e.g. acquisitive for aboveground -high SLA- and conservative for belowground- low SRL). Positive co-variation or lack of correlation between aboveground and belowground strategies were interpreted as reflecting the absence of constrains.

Patterns of community assembly for each trait were investigated with the functional richness index (Fric; sensu Villéger et al. 2008). This index has been previously linked to community assembly processes and used for detecting processes of habitat filtering and limiting similarity (e.g. driven by competition; Cornwell et al. 2006, Mason et al. 2013, Botta-Dukát and Czúcz 2016). Fric relies on the convex hull concept (i.e. the minimum convex hull which includes all species considered; Cornwell et al. 2006). Variation in Fric should reflect a variation in the range of functional strategies within the plant community, with low and high Fric pointing to, respectively, habitat filtering (niche overlapping) and limiting similarity (niche partitioning) processes in action (Cornwell et al. 2006, Mason et al. 2013). However, since Fric may increase as a

function of species richness (Mason et al. 2013), and detecting community assembly processes requires comparing observed diversity with that expected under a null model, we calculated the standardized-effect-size of Fric (SES-Fric), as follows:

$$SES(Fric) = \frac{Observed Fric - Mean Expected Fric}{SD Expected Fric}$$
 (2)

When SES-Fric < 0, the observed functional richness in a community is lower than expected by chance ('functional convergence' or 'niche overlapping'). Instead, when SES-Fric > 0, the observed functional richness is higher than expected by chance ('functional divergence' or 'niche partitioning'). We calculated expected Fric values by shuffling trait values across all species within the dataset 999 times. This algorithm is equally sensitive to functional convergence due to habitat filtering as well as to functional divergence due to limiting similarity processes (Botta-Dukát and Czúcz 2016). To assess the variation of functional patterns along the sea-inland stress gradient, for each trait we computed Pearson's coefficient correlation between the SES-Fric values and sea distance. As for CWMs, we concluded on the presence of a trade-off (limiting similarity processes acting between the aboveground and belowground functional traits) if e.g. in plant communities aboveground strategies converged (SES-Fric < 0) and belowground traits diverged (SES-Fric > 0). Positive co-variation or its absence represents evidence for independence, as this would indicate that assembly of species via aboveground traits is not constrained by belowground functional patterns.

All analyses were run in the R environment (ver. 4.1.2, www.r-project.org). Functional richness (Fric), CWM and CM

were calculated with the *dbFD* function in the 'FD' package (Laliberté et al. 2014), while the fourth corner analysis was run with *test fourth* function in the 'weimea' package (Zelený 2020).

#### **Results**

Regarding the variation in the functional strategies (CWM), we found significant and positive co-variation with distance to the sea for SLA (corr=0.18;Fig. 2c) and BOWC (corr=0.14;Fig. 2f) and significant negative co-variation for SRL (corr= -0.19; Fig. 2e). Moreover, we found co-variation close to significant for the SM (corr=-0.13;Fig. 2c) and mostly a lack of co-variation for both H (corr=-0.12; Fig. 2a) and RTD (corr=0.06;Fig. 2d). Similar patterns resulted also from the unweighted version of this index (i.e. community mean, CM), with the only exception of the negative co-variation of H and SM which became significant (Supporting information).

Concerning community assembly patterns (SES-Fric), we found a significant variation in the range of functional strategies (SES-Fric) along the sea-inland environmental

gradient for all the considered traits (Fig. 3). There was a similar positive trend of all the traits with sea distance, ranging from a convergence pattern (SES-Fric < 0) in plant communities closer to the sea, to a divergence pattern (SES-Fric > 0) in more inland plant communities. The strongest positive correlation resulted for the aboveground traits of SLA and SM (corr=0.31 and corr=0.39, respectively; Fig. 3b-c) and for the belowground traits SRL and BOWC (corr=0.36 and corr=0.32, respectively; Fig. 3e-f).

#### **Discussion**

In this study we found that plant trait variation followed partly predictions from the resource allocation theory but was also partly independent for above and belowground traits, depending on which aspect of community organization was involved. Specifically, the variation in the functional strategies (CWM or CM) confirmed our first scenario, i.e. the presence of a tradeoff in the resource allocation between aboveground and belowground organs (Tilman 2020). Plant communities growing in stressful conditions (i.e. close to the seashore) were characterized

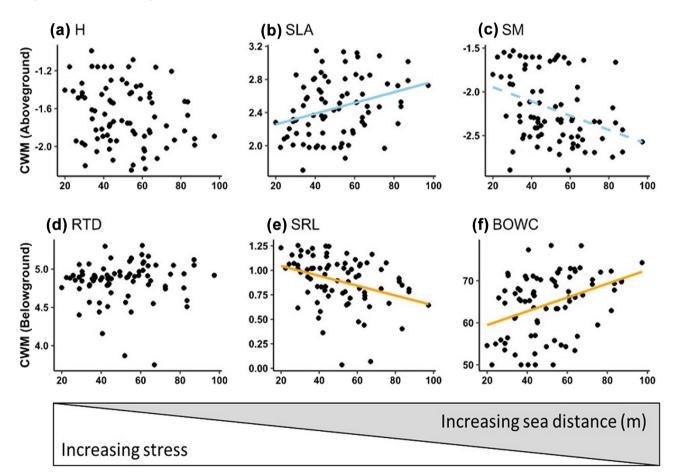


Figure 2. Co-variation of community weighted mean (CWM) in relation to sea distance (m) for aboveground traits: (a) plant height (H, m), (b) specific leaf area (SLA, mm<sup>2</sup> g<sup>-1</sup>), (c) seed mass (SM, g), and belowground traits: (d) root tissue density (RTD, mg cm<sup>-3</sup>), (e) specific root length (SRL, cm mg<sup>-1</sup>), (f) belowground organ water content (BOWC, %). Significant co-variations (p < 0.05) are represented by a solid skyblue and orange line for aboveground and belowground traits, respectively. Close to significant co-variation (p < 0.10) are represented by dotted line. All the selected traits, excepted for BOWC, were  $\log_{10}$ -transformed before calculating the CWM.

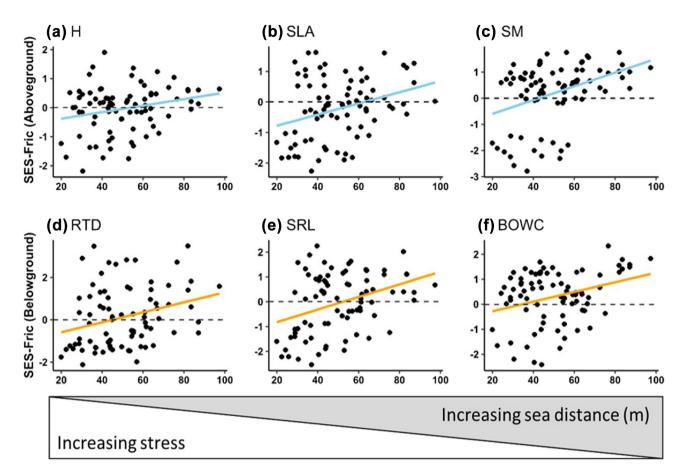


Figure 3. Co-variation of standardized effect size of functional richness (SES-Fric) in relation to sea distance (m) for aboveground traits: (a) plant height (H, m), (b) specific leaf area (SLA, mm $^2$  g $^{-1}$ ), (c) seed mass (SM, g), and belowground traits: (d) root tissue density (RTD, mg cm $^{-3}$ ), (e) specific root length (SRL, cm mg $^{-1}$ ), (f) belowground organ water content (BOWC, %). Significant co-variations (p < 0.05) are represented by a solid skyblue and orange line for aboveground and belowground traits, respectively. All the selected traits, excepted for BOWC, were log $_{10}$ -transformed before calculating the SES-Fric.

mainly by aboveground conservative strategies (low SLA) coupled with conservative water strategies (low BOWC) but with acquisitive fine-root strategies (high SRL). At the opposite end of the gradient, in less stressful conditions (i.e. far from the shoreline), plant communities showed high photosynthetic rate (high SLA) coupled with less conservative water-use (high BOWC) but high conservative fine-root strategies (low SRL) (Fig. 4a). In terms of assembly rules (SES-Fric), the results were in line with the second scenario, in which aboveground and belowground traits contribute independently to community assembly. Higher stress conditions promote functional convergence for all selected traits, probably driven by habitat filtering processes, while under lower stress conditions, functional divergence dominates for both above and belowground traits, consistently with strong limiting similarity processes, potentially related to more competitive interactions (Fig. 4b).

# Dominant strategies (CWM): resource allocation theory

Our observed variation in functional strategies supported Tilman's hypothesis (2020), that plant species occurring in

resource-rich environments should show acquisitive strategies mainly for aboveground organs, while belowground acquisitive strategies should be favored under stressful conditions. In our study, plant communities closer to the sea (foredunes) are subject to high stress (incoherent and unfertile soils, wind erosion, aridity and salinity) where water and soil nutrients represent the main limiting resource (Carboni et al. 2011). These conditions favor persistence of plant communities dominated mainly by taller species with long lifespan (Supporting information) like C. arundinacea and T. junceum, displaying overall aboveground conservative strategies complemented by water-conservative strategies (low SLA and low BOWC) (Stanisci et al. 2010). Indeed, species growing in drought conditions tend to use water more efficiently (Freschet et al. 2010, de Bello et al. 2012). Moreover, this environmental severity is also well reflected by the acquisitive pattern we found for fine-roots. In line with studies finding an increase of SRL in response to nutrient limitation and water availability, we found species growing closer to the sea are characterized by high SRL values (Zangaro et al. 2008, Holdaway et al. 2011, Kramer-Walter et al. 2016, De Battisti et al. 2020). High SRL is associated with enhanced

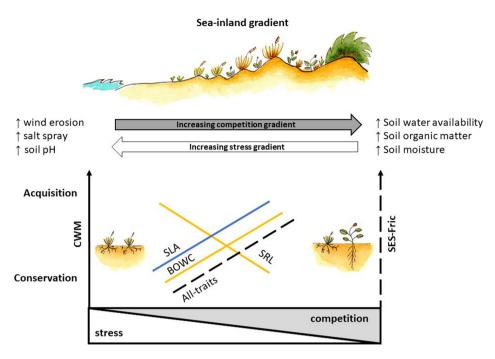


Figure 4. Conceptual illustration of the results showing different patterns in aboveground and belowground traits in relation to the seainland gradient as a proxy of a stress gradient. Our results indicate that functional dominant strategies (CWM) changed as predicted by the resource allocation theory with an increase towards stressful conditions of aboveground conservative (low SLA) and water-saving strategies (low BOWC) coupled with belowground acquisitive strategies (high SRL). However, all the selected traits showed an independent contribution (H, plant height; SLA, specific leaf area; SM, seed mass; SRL, specific root length; RTD, root tissue density; BOWC, belowground organ water content) to community assembly (SES-Fric) since they are all filtered by stressful conditions.

root proliferation and the production of thinner and lighter fine-roots, which in nutrient-poor environments represents an effective strategy to promote nutrient acquisition by exploring higher soil volume per unit carbon investment in root length (Laliberté et al. 2015).

Moving away from the sea (dune grasslands), the functional portfolio of plant communities shifts completely. Here, where stress relaxes and resource availability increases, plant communities showed aboveground acquisitive strategies (high SLA) coupled with less conservative water-use strategies (high BOWC). This pattern is perfectly in line with the trends we found belowground for SRL, which was lower in grasslands far from the sea with more coherent and nutrientrich soils. Indeed, it has been suggested that under fertile soils thicker root cortex (i.e. low SRL) supports higher metabolic rates (Chimungu et al. 2014, Kramer-Walter et al. 2016). This functional portfolio of plant communities far away from the sea can be also ascribed to the increase of annual species compared to foredunes (Supporting information). Generally, annual species have shorter lifespans, investing most of the energy in fast growth (high SLA and SRL) and in considerable seed production, resulting in less conservative water-use (high BOWC) (Supporting information). However, contrary to what reported in the literature (Roumet et al. 2006), here we found annual species with lower SRL compared to non annual ones (Supporting information), probably as a consequence of a lower root proliferation in the soil. These shifts in the dominant functional strategies were confirmed

and emerged even more clearly when considering presence/ absence species data (CM, Supporting information). Variation in CWMs could result just from changes in the cover of few dominant species, while variation in CM exclusively depends on species turnover (de Bello et al. 2021), confirming that the considerable turnover in species composition due to the sea-inland gradient also reflects in a clear functional turnover (Acosta et al. 2003). Finally, we should note that in our case, weighting belowground traits with aboveground species cover or using the unweighted version of the index, produce consistent results. However, we cannot exclude that incorporating abundances measured belowground could produce different results, calling for further investigation (Tsakalos et al. 2022).

## Community assembly patterns (SES-Fric): independent contribution

The existence of tradeoffs between functional dimension in community assembly processes is a fascinating idea, yet only partially supported by evidence (Lhotsky et al. 2016, Vojtkó et al. 2017). Given that the overlap of species niches within a community is limited, it is expected that if some traits converge, others should diverge and that this opposite pattern should be stronger in case of traits capturing similar functions for different functional dimension (i.e. SLA for aboveground and SRL for belowground) (MacArthur and Levins 1967, Lhotsky et al. 2016, Vojtkó et al. 2017). However, our study challenges the generality of this tradeoff, at least in terms of

LHS aboveground traits vs. belowground organ and fine-root traits. We found significant variation for all traits towards the same direction (i.e. functional convergence under higher stress conditions), independently of whether these were above or belowground traits. The functional convergence can be ascribed to the effect of habitat filtering, as only specialized species are able to persist in such harsh ecological conditions (Ricotta et al. 2012, Carboni et al. 2016).

While functional convergence for aboveground traits confirms previous results from these environments (Ricotta et al. 2012, Carboni et al. 2016), the belowground functional pattern does not support the presence of above-belowground tradeoff. Therefore, an interesting question arises: how can species coexist in the stressful conditions close to the seashore if they are functionally similar both in their aboveground and belowground traits? One possible explanation is that traits may have diverse competitive outcomes affecting mechanisms of species coexistence. For example, belowground traits like SRL and BOWC should connect to sizesymmetric competition (Rajaniemi 2003, Mayfield and Levine 2010), meaning that small differences in competitive ability do not have disproportionate effects on competitive exclusion. Thus, species with similar niches are more likely to coexist (Rajaniemi 2003). Additionally, selective pressure for niche differentiation between co-existing species should be less intense in stressed communities and fitness should be guaranteed by traits that maximize acquisition and retention of limiting belowground resources (Holdaway et al. 2011, Mason et al. 2013).

In contrast, other traits like plant height are related to size-asymmetric competition, which means that a small difference in plant size may promote strong competitive exclusion (Rajaniemi 2003). Taller species displace shorter ones and the resulting plant community is composed only by dominant species displaying similar size (Lepš 2014). However, plant communities close to the sea are characterized by scattered vegetation with very low cover, suggesting that competitive exclusion is not a key community assembly driver (Santoro et al. 2012).

At the opposite end of the gradient, i.e. far away from the sea, the presence of higher resource availability promotes functional divergence for all considered traits. Our finding supports the idea that resource-rich communities have greater functional diversity than resource-poor communities both above and belowground (Mouchet et al. 2010, Mason et al. 2013). This functional pattern is consistent with strong limiting similarity, a hypothesis suggesting that species that are quite different can assemble even under strong competition because they exploit different niches (Stubbs and Wilson 2004, Carboni et al. 2016). Such higher diversity is also consistent with the presence of species having completely different life-histories, including annual (e.g. S. canescens, F. fasciculata) and non-annual species (e.g. Lotus cytisoides, Plantago macrorrhiza) (Supporting information). Indeed, a functional divergence is expected when species belonging to different life lifespan coexist together (Gross et al. 2013, Klimešová and Herben 2023).

#### Conclusion

Coastal dunes are very peculiar ecosystems characterized by a strong sea-inland gradient of stress which determines a tradeoff in resource allocation between above- and belowground organs. Nonetheless, belowground traits assembled independently of aboveground traits, providing poor support for the idea that limiting similarity results in the differentiation of belowground traits to compensate for functional convergence of aboveground traits. Close to the sea, low nutrient and water availability limits the number of species able to survive and filters species with a stress-tolerant strategy, i.e. mostly perennial species with water-saving strategies and acquisitive root system aimed at a better soil exploration of the limiting water/nutrient resource. Along the gradient, moving away from the sea stress, the increase in nutrients and soil moisture allows more diverse species to survive by partitioning the available niche space. As stress conditions relax, we observed the coexistence of both stress-avoidance (i.e. annual species) and stress-tolerant strategies (i.e. perennial species), displaying overall higher aboveground acquisitive strategies but conservative fine-root traits. We are aware that other functional dimensions can play a role in promoting species coexistence like phenology, however the selected traits probably represent those explaining higher variation in the plant functional spectrum. Our study reinforces the viewpoint that community organization is more complex than expected and, more importantly, that the exclusive use of aboveground traits provides only a partial picture of plant communities.

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Alessandro Bricca: Conceptualization (lead); Formal analysis (lead); Investigation (lead); Methodology (lead); Visualization (lead); Writing – original draft (lead); Writing – review and editing (lead). Marta G. Sperandii: Formal analysis (supporting); Writing – review and editing (supporting). Alicia T. R. Acosta: Data curation (lead); Visualization (supporting); Writing – review and editing (supporting). Antonio Montagnoli: Conceptualization (supporting); Methodology (supporting); Software (supporting); Writing

review and editing (supporting). Greta La Bella: Writing – review and editing (supporting). Mattia Terzaghi: Software (lead). Marta Carboni: Conceptualization (supporting); Visualization (supporting); Writing – original draft (supporting); Writing – review and editing (supporting).

#### Data availability statement

Data are available from the Dryad Digital Repository: https://doi.org/doi:10.5061/dryad.d51c5b07r (Bricca et al. 2023b).

#### **Supporting information**

The Supporting information associated with this article is available with the online version.

#### References

- Acosta, A., Stanisci, A., Ercole, S. and Blasi, C. 2003. Sandy coastal landscape of the Lazio region (central Italy). Phytocoenologia 33: 715–726.
- Acosta, A., Izzi, C. F. and Stanisci, A. 2006. Comparison of native and alien plant traits in Mediterranean coastal dunes. – Commun. Ecol. 7: 35–41.
- Acosta, A., Ercole, S., Stanisci, A., Pillar, V. D. P. and Blasi, C. 2007. Coastal vegetation zonation and dune morphology in some mediterranean ecosystems. J. Coast. Res. 23: 1518–1524.
- Acosta, A., Carranza, M. L. and Izzi, C. F. 2009. Are there habitats that contribute best to plant species diversity in coastal dunes?. – Biodivers. Conserv. 18: 1087–1098.
- Bazzichetto, M., Malavasi, M., Acosta, A. T. R. and Carranza, M. L. 2016. How does dune morphology shape coastal EC habitats occurrence? A remote sensing approach using airborne LiDAR on the Mediterranean coast. Ecol. Indic. 71: 618–626.
- Botta-Dukát, Z. and Czúcz, B. 2016. Testing the ability of functional diversity indices to detect trait convergence and divergence using individual-based simulation. Methods Ecol. Evol. 7: 114–126.
- Bricca, A., Catorci, A. and Tardella, F. M. 2020. Intra-specific multi-trait approach reveals scarce ability in the variation of resource exploitation strategies for a dominant tall-grass under intense disturbance. – Flora 270: 151665.
- Bricca, A., Tardella, F. M., Ferrara, A., Xinfang, X., Tolu, F. and Catorci, A. 2021. Environmental heterogeneity compensates the potential homogenising effect of abandonment of grazing in a sub-Mediterranean mountain landscape. Plant Ecol. Divers. 14: 223–243.
- Bricca, A., Di Musciano, M., Ferrara, A., Theurillat, J. P. and Cutini, M. 2022. Community assembly along climatic gradient: contrasting pattern between-and within-species. – Perspect. Plant Ecol. 56: 125675.
- Bricca, A., Bonari, G., Cubino, J. P. and Cutini, M. 2023a. Effect of forest structure and management on the functional diversity and composition of understory plant communities. – Appl. Veg. Sci. 26: e12710.
- Bricca, A., Sperandii, M. G., Acosta, A. T. R., Montagnoli, A., Bella, G. L., Terzaghi, M., Carboni, M. 2023b. Data from: Above- and belowground traits along a stress gradient: trade-off or not? .

   Dryad Digital Repository, https://doi.org/doi:10.5061/dryad.d51c5b07r.

- Carboni, M., Santoro, R. and Acosta, A. T. R. 2011. Dealing with scarce data to understand how environmental gradients and propagule pressure shape fine-scale alien distribution patterns on coastal dunes. J. Veg. Sci. 22: 751–765.
- Carboni, M., de Bello, F., Janeček, Š., Doležal, J., Horník, J., Lepš, L., Reitalu, T. and Klimešová, J. 2014. Changes in trait divergence and convergence along a productivity gradient in wet meadows. Agric. Ecosyst. Environ. 182: 96–105.
- Carboni, M., Zelený, D. and Acosta, A. T. R. 2016. Measuring ecological specialization along a natural stress gradient using a set of complementary niche breadth indices. J. Veg. Sci. 27: 892–903.
- Carmona, C. P., Bueno, C. G., Toussaint, A., Träger, S., Díaz, S., Moora, M., Munson, A. D., Pärtel, M., Zobel, M. and Tamme, R. 2021. Fine-root traits in the global spectrum of plant form and function. – Nature 597: 683–687.
- Chimungu, J. G. Brown, K. M. and Lynch, J. P. 2014. Reduced root cortical cell file number improves drought tolerance in maize. Plant Physiol. 166: 1943–1955.
- Ciccarelli, D. 2015. Mediterranean coastal dune vegetation: are disturbance and stress the key selective forces that drive the psammophilous succession?. Estuar. Coast. Shelf Sci. 165: 247–253.
- Ciccarelli, D. and Bona, C. 2022. Exploring the functional strategies adopted by coastal plants along an ecological gradient using morpho-functional traits. Estuar. Coast. 45: 114–129.
- Coomes, D. A. and Grubb, P. J. 2000. Impacts of root competition in forests and woodlands: a theoretical framework and review of experiments. Ecol. Monogr. 70: 171–207.
- Cornwell, W. K., Schwilk, D. W. and Ackerly, D. D. 2006. A traitbased test for habitat filtering: convex hull volume. – Ecology 87: 1465–1471.
- Coyle, J. R., Halliday, F. W., Lopez, B. E., Palmquist, K. A., Wilfahrt, P. A. and Hurlbert, A. H. 2014. Using trait and phylogenetic diversity to evaluate the generality of the stress-dominance hypothesis in eastern North American tree communities. Ecography 37: 814–826.
- De Battisti, D., Fowler, M. S., Jenkins, S. R., Skov, M. W., Bouma, T. J., Neyland, P. J. and Griffin, J. N. 2020. Multiple trait dimensions mediate stress gradient effects on plant biomass allocation, with implications for coastal ecosystem services. J. Ecol. 108: 1227–1240.
- de Bello, F., Janeček, Š., Lepš, J., Doležal, J., Macková, J., Lanta, V. and Klimešová, J. 2012. Different plant trait scaling in dry versus wet central European meadows. J. Veg. Sci. 23: 709–720.
- de Bello, F., Carmona, C. P., Dias, A. T., Götzenberger, L., Moretti, M. and Berg, M. P. 2021. Handbook of trait-based ecology: from theory to R tools. – Cambridge Univ. Press.
- de la Riva, E. G., Tosto, A., Pérez-Ramos, I. M., Navarro-Fernández, C. M., Olmo, M., Anten, N. P., Marañón, T. and Villar, R. 2016. A plant economics spectrum in Mediterranean forests along environmental gradients: is there coordination among leaf, stem and root traits?. J. Veg. Sci. 27: 187–199.
- de la Riva, E. G., Godoy, O., Castro-Díez, P., Gutiérrez-Cánovas, C. and Vilà, M. 2019. Functional and phylogenetic consequences of plant invasion for coastal native communities. J. Veg. Sci. 30: 510–520.
- Díaz, S. et al. 2016. The global spectrum of plant form and function. Nature 529: 167–171.
- Erktan, A., McCormack, M. L. and Roumet, C. 2018. Frontiers in root ecology: recent advances and future challenges. – Plant Soil 424: 1–9.

- Freschet, G. T., Cornelissen, J. H., Van Logtestijn, R. S. and Aerts, R. 2010. Evidence of the 'plant economics spectrum'in a subarctic flora. - J. Ecol. 98: 362-373.
- Funk, J. L., Larson, J. E., Ames, G. M., Butterfield, B. J., Cavender-Bares, J., Firn, J., Laughlin, D. C., Sutton-Grier, A. E., Williams, L. and Wright, J. 2017. Revisiting the Holy Grail: using plant functional traits to understand ecological processes. -Biol. Rev. 92: 1156-1173.
- Garnier, E., Cortez, J., Billès, G., Navas, M. L., Roumet, C., Debussche, M., Laurent, G., Blanchard, A., Aubry, D., Bellmann, A., Neill, C. and Toussaint, J-P. 2004. Plant functional markers capture ecosystem properties during secondary succession. - Ecology 85: 2630-2637.
- Götzenberger, L., de Bello F, Bråthen KA, Davison J, Dubuis A, Guisan, A., Lepš, J., Lindborg, R., Moora, M., Pärtel, M., Pellissier, L., Pottier, J., Vittoz, P., Zobel, K. and Zobel, M. 2012. Ecological assembly rules in plant communities-approaches, patterns and prospects. - Biol. Rev. 87:111-127.
- Gross, N., Börger, L., Soriano-Morales, S. I., Le Bagousse-Pinguet, Y., Quero, J. L., García-Gómez, M., Valencia-Gómez, E. and Maestre., F. T. 2013. Uncovering multiscale effects of aridity and biotic interactions on the functional structure of Mediterranean shrublands. - J. Ecol. 101: 637-649.
- Holdaway, R. J., Richardson, S. J., Dickie, I. A., Peltzer, D. A. and Coomes, D. A. 2011. Species- and community-level patterns in fine root traits along a 120000- year soil chronosequence in temperate rain forest. – J. Ecol. 99: 954–963. Kattge, J. et al. 2020. TRY plant trait database–enhanced coverage
- and open access. Global Change Biol. 26: 119-188.
- Klimešová, J. and Herben, T. 2023. The hidden half of the fine root differentiation in herbs: nonacquisitive belowground organs determine fine-root traits. - Oikos 2023: e08794.
- Kramer-Walter, K. R., Bellingham, P. J., Millar, T. R., Smissen, R. D., Richardson, S. J. and Laughlin, D. C. 2016. Root traits are multidimensional: specific root length is independent from root tissue density and the plant economic spectrum. - J. Ecol. 104: 1299-1310.
- Laliberté, E. 2017. Below-ground frontiers in trait-based plant ecology. - New Phytol. 213: 1597-1603.
- Laliberté, E., Legendre, P. and Shipley, B. 2014. FD: measuring functional diversity from multiple traits, and other tools for functional ecology. - R package ver. 1.0-12.1.
- Laliberté, E., Lambers, H., Burgess, T. I. and Wright, S. J. 2015. Phosphorus limitation, soil-borne pathogens and the coexistence of plant species in hyperdiverse forests and shrublands. -New Phytol. 206: 507-521.
- Laughlin, D. C., Mommer, L., Sabatini, F. M., Bruelheide, H., Kuyper, T. W., McCormack, M. L., Bergmann, J., Freschet, G. T., et al. 2021. Root traits explain plant species distributions along climatic gradients yet challenge the nature of ecological tradeoffs. - Nat. Ecol. Evol. 5: 1123-1134.
- Lepš, J. 2014. Scale-and time-dependent effects of fertilization, mowing and dominant removal on a grassland community during a 15-year experiment. – J. Appl. Ecol. 51: 978–987.
- Lhotsky, B., Kovács, B., Ónodi, G., Csecserits, A., Rédei, T., Lengyel, A., Kertész, M. and Botta-Dukát, Z. 2016. Changes in assembly rules along a stress gradient from open dry grasslands to wetlands. - J. Ecol. 104: 507-517.
- Lubbe, F. C., Bitomský, M., Hajek, T., de Bello, F., Doležal, J., Jandová, V., Janeček, Š., Bartušková, A., Lanta, V. and Klimešová, J. 2021. A tale of two grasslands: how belowground

- storage organs coordinate their traits with water-use traits. -Plant Soil 465: 533-548.
- MacArthur, R. and Levins, R. 1967. The limiting similarity, convergence, and divergence of coexisting species. – Am. Nat. 101: 377-385.
- Májeková, M., Paal, T., Plowman, N. S., Bryndová, M., Kasari, L., Norberg, A., Weiss, M., Bishop, T. R., Luke, S. H., Sam, K., Le Bagousse-Pinguet, Y., Lepš, J., Götzenberger, L. and de Bello, F. 2016. Evaluating functional diversity: missing trait data and the importance of species abundance structure and data transformation. - PloS One 11: e0149270.
- Mason, N. W., de Bello, F., Mouillot, D., Pavoine, S. and Dray, S. 2013. A guide for using functional diversity indices to reveal changes in assembly processes along ecological gradients. - J. Veg. Sci. 24: 794-806.
- Mayfield, M. M. and Levine, J. M. 2010. Opposing effects of competitive exclusion on the phylogenetic structure of communities. - Ecol. Lett. 13: 1085-1093.
- Montagnoli, A., Terzaghi, M., Di Iorio, A., Scippa, G. S. and Chiatante, D. 2012. Fine-root morphological and growth traits in a Turkey-oak stand in relation to seasonal changes in soil moisture in the southern Apennines, Italy. – Ecol. Res. 27:1015–1025.
- Montagnoli, A., Di Iorio, A., Terzaghi, M., Trupiano, D., Scippa, G. S. and Chiatante, D. 2014. Influence of soil temperature and water content on fine root seasonal growth of European beech natural forest in southern Alps, Italy. - Eur. J. For. Res. 133:957-968.
- Mouchet, M. A., Villeger, S., Mason, N. W. H. and Mouillot, D. 2010. Functional diversity measures: an overview of their redundancy and their ability to discriminate community assembly rules. - Funct. Ecol. 24: 867-876.
- Pakeman, R. J. and Quested, H. M. 2007. Sampling plant functional traits: what proportion of the species need to be measured?. - Appl. Veg. Sci. 10: 91-96.
- Peres-Neto, P. R., Dray, S. and ter Braak, C. J. 2017. Linking trait variation to the environment: critical issues with communityweighted mean correlation resolved by the fourth-corner approach. - Ecography 40: 806-816.
- Perez-Harguindeguy, N. et al. 2013. Corrigendum to: new handbook for standardised measurement of plant functional traits worldwide. - Aust. Bot. 64: 715-716.
- Rajaniemi, T. K. 2003. Explaining productivity-diversity relationships in plants. - Oikos 101: 449-457.
- Reynolds, H. L. and D'antonio, C. 1996. The ecological significance of plasticity in root weight ratio in response to nitrogen: opinion. – Plant Soil 185: 75–97.
- Ricotta, C., Pavoine, S., Bacaro, G. and Acosta, A. T. R. 2012. Functional rarefaction for species abundance data. - Methods Ecol. Evol. 3: 519-525.
- Roumet, C., Urcelay, C. and Díaz, S. 2006. Suites of root traits differ between annual and perennial species growing in the field. New phytol. 170: 357–368.
- Santoro, R., Jucker, T., Carboni, M. and Acosta, A. T. R. 2012. Patterns of plant community assembly in invaded and noninvaded communities along a natural environmental gradient. J. Veg. Sci. 23: 483–494.
- Sperandii, M. G., Prisco, I., Stanisci, A. and Acosta, A. T. R. 2017. RanVegDunes-a random plot database of Italian coastal dunes. Phytocoenologia 47: 231–232.
- Spitzer, C. M., Sundqvist, M. K., Wardle, D. A., Gundale, M. J. and Kardol, P. 2022. Root trait variation along a sub-arctic tundra elevational gradient. - Oikos 2023: e08903.

- Stanisci, A., Acosta, A. T. R., Di Iorio, A. and Vergalito, M. 2010. Leaf and root trait variability of alien and native species along Adriatic coastal dunes (Italy). – Plant Biosyst. 144: 47–52.
- Stubbs, W. J. and Wilson, B. J. 2004. Evidence for limiting similarity in a sand dune community. J. Ecol. 92: 557–567.
- ter Braak, C. J. F., Peres-Neto, P. and Dray, S. 2018. Simple parametric tests for trait—environment association. J. Veg. Sci. 29: 801–811.
- Tilman, D. 2020. Plant strategies and the dynamics and structure of plant communities, (MPB-26), vol. 26. Princeton Univ. Press.
- Tsakalos, J. L., Ottaviani, G., Chelli, S., Rea, A., Elder, S., Dobrowolski, M. P. and Mucina, L. 2022. Plant clonality in a soil-impoverished open ecosystem: insights from southwest Australian shrublands. Ann. Bot. 130: 981–990.
- Van Der Maarel, E. 2003. Some remarks on the functions of European coastal ecosystems. Phytocoenologia 33: 187–202.
- Villéger, S., Mason, N. W. and Mouillot, D. 2008. New multidimensional functional diversity indices for a multifaceted framework in functional ecology. – Ecology 89: 2290–2301.

- Vojtkó, A., Freitag, M., Bricca, A., Martello, F., Compañ, J. M., Küttim, M., Kun, R., de Bello, F., Klimešová, J. and Götzenberger, L. 2017. Clonal vs leaf-height-seed (LHS) traits: which are filtered more strongly across habitats?. Folia Geobot. 52: 269–281.
- Weiher, E. and Keddy, P. A. 1995. Assembly rules, null models, and trait dispersion: new questions from old patterns. Oikos 74: 159–164.
- Westoby, M. 1998. A leaf-height-seed (LHS) plant ecology strategy scheme. Plant Soil 199: 213–227.
- Zangaro, W., de Assis, R. L., Rostirola, L. V., Souza, P., Goncalves, M. C., Andrade, G. and Nogueira, M. A. 2008. Changes in arbuscular mycorrhizal associations and fine root traits in sites under different plant successional phases in southern Brazil. Mycorrhiza 19: 37–45.
- Zelený, D. 2018. Which results of the standard test for community-weighted mean approach are too optimistic?. J. Veg. Sci. 29: 953–966.
- Zelený, D. 2020. weimea: weighted mean analysis. R package ver. 0.1.18.